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ORIGINAL ARTICLE

Effects of vegetation type on microbial biomass carbon and nitrogen in subalpine mountain forest soils



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Received 7 April 2013; received in revised form 8 January 2014; accepted 10 February 2014

Available online 21 March 2014

KEYWORDS

Forest vegetation;
Microbial biomass
carbon and
nitrogen;
Microbial populations

Background/purpose: Microbial biomass plays an important role in nutrient transformation and conservation of forest and grassland ecosystems. The objective of this study was to determine the microbial biomass among three vegetation types in subalpine mountain forest soils of Taiwan.

Methods: Tachia is a typical high-altitude subalpine temperate forest ecosystem in Taiwan with an elevation of 1800–3952 m and consists of three vegetation types: spruce, hemlock, and grassland. Three plots were selected in each vegetation type. Soil samples were collected from the organic layer, topsoil, and subsoil. Microbial biomass carbon (C_{mic}) was determined by the chloroform fumigation–extraction method, and microbial biomass nitrogen (N_{mic}) was determined from the total nitrogen (N_{tot}) released during fumigation–extraction. Bacteria, actinomycetes, fungi, cellulolytic microbes, phosphate-solubilizing microbes, and nitrogen-fixing microbes were also counted.

Results: The C_{mic} and N_{mic} were highest in the surface soil and declined with the soil depth. These were also highest in spruce soils, followed by in hemlock soils, and were lowest in grassland soils. C_{mic} and N_{mic} had the highest values in the spring season and the lowest values in the winter season. C_{mic} and N_{mic} had significantly positive correlations with total organic carbon (C_{org}) and N_{tot} . Contributions of C_{mic} and N_{mic} , respectively, to C_{org} and N_{tot} indicated that the microbial biomass was immobilized more in spruce and hemlock soils than in grassland soils. Microbial populations of the tested vegetation types decreased with increasing soil depth.

Conclusion: C_{mic} and N_{mic} were high in the organic layer and decreased with the depth of layers. These values were higher for spruce and hemlock soils than for grassland soils. Positive correlations were observed between C_{mic} and N_{mic} and between C_{org} and N_{tot} .

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Introduction

Natural forests have gained recognition as the sites with high biodiversity. The complex relationships among fauna, flora, and microflora are due to the richness of habitat. Microbial biomass comprises about 2–3% of total organic carbon (C_{org}) and represents a major labile pool of nutrients in the soil.^{1,2} It plays an important role in nutrient transformation and conservation of forest and grassland ecosystems, in both tropical and temperate climates.^{3–8} The microbial biomass is influenced by soil organic matter content,^{9,10} temperature,^{4,11} moisture content,¹² and pH.¹³ Soil microbial biomass is an important indicator of soil fertility in soil ecological studies^{14,15} and sustainable environmental management.¹⁶ The ratio of microbial biomass carbon to total organic carbon (C_{mic}/C_{org}) in soil may serve as a quantitative indicator of soil carbon dynamics.^{17,18}

Forest vegetation affects the microbial processes of carbon and nitrogen cycles due to the differences in quality and quantity of litters, root exudates, and soil properties.^{11,19–21} Tree species has an impact on soil fertility and microbial community composition, which in turn can affect the soil microbial biomass and microbial efficiency in carbon utilization.^{3,9}

Microbial biomass varies with the seasonal patterns of soil temperature, moisture content, and substrate availability,^{22,23} and also with seasons and soil depth.^{24,25} Sarathchandra et al.²² showed that soil microbial biomass was greater in spring and autumn than in summer and winter; by contrast, Kaiser and Heinemeyer²⁶ observed higher microbial biomass in summer than in winter. Changes of microbial biomass with soil depth had been well documented.^{24,25}

The size of soil microbial biomass is usually resource limitation, and the C_{mic} and microbial biomass nitrogen (N_{mic}) are generally related to the soil organic matter content.^{4,10} The different chemical compositions of roots and leaves can affect the compositions and functions of the microbial community. The amount of information on soil microbial biomass under different vegetation types in forest ecosystems is limited. Therefore, the objective of this study was to determine the microbial biomass under the three vegetation types in subalpine mountain forest soils of Taiwan. Seasonal variations of microbial biomass and microbial populations were also investigated.

Materials and methods

Site description and soil sampling

This study was conducted at Tatachia forest, in the saddle of Jade Mountain, central Taiwan (120° 52' E, 23° 28' N). Tatachia is a typical high-altitude subalpine temperate forest ecosystem in Taiwan with an elevation of 1800–3952 m. Tatachia forest (>400 ha) has been identified, by the Taiwan Long Term Ecological Study Network, as one of the four natural forest sites of subalpine forest ecosystems to study the long-term monitoring of environmental variables, plant successions, ecosystem phenomena and processes, dynamics of animal communities, soil nutrient movements, human impacts (recreational activities), ecological modeling, and data management. The study

area geologically consists of metamorphosed sedimentary rock (*Miocene epoch*) comprising sandstone and shale.

Three permanent plots, each for a major vegetation type, were established in Tatachia subalpine ecosystem. The hemlock area is a thick forest floor dominated by Chinese hemlock (*Tsuga chinensis*) and dwarf bamboo (*Yushania nikitayamensis*). The spruce area is a deep soil ecosystem dominated by spruce (*Picea morrisonicola*). Less dominant species include Taiwan false cypress (*Chamaecyparis formosensis*) and Armands pine (*Pinus armandi*). The grassland area is dominated by alpine silver grass (*Miscanthus trans-morrisonensis*). The spruce and hemlock areas have the greatest canopy patches of woody plants, whereas the grassland area has the greatest canopy gaps and the shallowest soil formation due to erosion. Sampling was done after the selection of representative plots under three vegetation types. In each vegetation type, three 2 m × 2 m plots were selected; three replicate soil samples were collected from each plot at the organic layer, topsoil, and subsoil by digging vertically from the surface. The organic layer was collected from 0–5 cm soil layer after removal of the above-ground plant debris. Soils underneath the organic layer were then sampled at topsoil (6–20 cm depth) and subsoil (21–40 cm depth). The soils were blackish–brown (organic layer), grey–blackish grey (topsoil), and grey–yellowish grey (subsoil) in color. Soil samples were collected during March–May, June–August, September–November, and December–February, representing spring, summer, autumn, and winter seasons, respectively. The sampling sites were the same in all sampling periods. The collected soils were packed separately in plastic bags and rapidly transported to the laboratory. Visible materials such as roots and litters were manually removed prior to sieving through a 2 mm sieve. Soil samples were stored at 4°C in the dark until use.

Microbial biomass carbon and nitrogen

The C_{mic} was determined by the chloroform fumigation–extraction method.²⁷ Fresh soil was adjusted to 55% of water-holding capacity and preincubated at 25°C for 7 days prior to the measurement. Soils were fumigated for 24 hours with alcohol-free chloroform ($CHCl_3$) vapors. After the fumigant was removed, soils were extracted with 0.5 M K_2SO_4 . The nonfumigated control soils were extracted under the same conditions at the time fumigation commenced. Fumigated soil extract (8.0 mL), 0.066 M $K_2Cr_2O_7$ (2 mL), HgO (70 mg), concentrated (conc.) H_2SO_4 (10.0 mL), and 85% H_3PO_4 (5.0 mL) were mixed thoroughly. The mixture was digested at 150°C for 30 minutes and titrated with 0.033 M ferrous (II) ammonium sulfate, using 1,10-phenanthroline–ferrous sulfate mixture as an indicator. The C_{mic} was calculated according to the method described by Wu et al.²⁸ $C_{mic} = E_C/K_{EC}$, where E_C is the difference between the C extracted from the fumigated and nonfumigated samples and $K_{EC} = 0.45$. The N_{mic} was determined by analyzing total nitrogen (N_{tot}) in the 0.5 M K_2SO_4 extract obtained from fumigation–extraction.²⁹ The N_{tot} of the extract was measured by the modified Kjeldahl method.³⁰ The N_{mic} was calculated following the method described by Brookes et al.²⁹ $N_{mic} = E_N/K_{EN}$, where E_N is the difference between the N extracted from the fumigated and nonfumigated samples and $K_{EN} = 0.54$.

Microbial populations

Bacteria were counted at 25°C for 5 days on nutrient agar (Merck, Darmstadt, Germany). Actinomycetes were cultivated at 25°C for 7 days on a glycerol–yeast extract medium consisting of (g/L) glycerol 5.0, yeast extract 2.0, K₂HPO₄ 1.0, and agar 15.0 at pH 7.0 ± 0.1. Streptomycin and cycloheximide were added to inhibit the growth of bacteria and fungi at a final concentration of 10 µg/mL.³¹ Fungi were grown at 25°C for 5 days on the Rose Bengal medium containing (g/L) glucose 10.0, peptone 5.0, K₂HPO₄ 1.0, MgSO₄·7H₂O 0.5, Rose Bengal 0.033, and agar 15.0 at pH 6.8 ± 0.1. Cellulolytic microbes were assayed at 25°C after 7 days of incubation on a modified Mandels–Reese medium with carboxymethylcellulose (Sigma-Aldrich, St. Louis, MO, USA) as the sole carbon source and sprayed with Congo red to show a clear zone around the colonies.³² Phosphate-solubilizing microbes were measured at 25°C after 5 days on a rock phosphate medium from the clear zone around the colonies.³³ Nitrogen-fixing microbes were counted after incubation at 25°C for 7 days on a nitrogen-free mannitol medium.²⁴ All experiments were carried out in triplicate.

Chemical analysis

Moisture content was determined by drying the sample overnight at 105°C to a constant weight. Soil pH was measured in five times volume of distilled water equilibrated with soil for 1 hour by a pH meter (Good digital pH meter model 2002, Plasma Equipment, Taiwan). Air temperatures were determined directly and under 5 cm depth of soil, respectively, with a thermometer. The N_{tot} was measured by the modified Kjeldahl method³⁰ and C_{org} was estimated by the modified Walkley–Black method, as described by Nelson and Sommers.³⁴ Soil sample (0.1 g) was mixed vigorously with 1 N K₂Cr₂O₇ (10 mL) and conc. H₂SO₄ (20 mL) for 30 minutes, and then 85% H₃PO₄ (10 mL) and distilled water (200 mL) were added to the mixture. After cooling, it was titrated with 0.5 N Fe(NH₄)₂(SO₄)₂ using diphenylamine as an indicator.

$$C_{\text{org}} (\%) = [(1 - S/B) \times 3.896] / \text{weight of soil sample} \times 100,$$

where *S* and *B* are the titer values of the soil and blank samples, respectively.

Statistical analysis

Analyses were carried out using triplicate samples, and the results were reported on a dry weight basis. Effects of vegetation type, soil horizon, and sampling date on soil properties were tested by the one-way analysis of variance and the Tukey multiple range tests using the statistical package Sigma Stat for Windows version 2.0.³⁵ Linear regression analyses were carried out to find out the relationship between C_{mic} and N_{mic} and other soil variables (C_{org}, N_{tot}, C/N ratio, pH, and moisture content), and it was considered significant at *p* < 0.05.

Results

Environmental conditions and soil properties

Mean air and soil temperatures of spruce, hemlock, and grassland vegetation types ranged from 5.40 ± 0.28°C to 20.72 ± 0.65°C and from 6.22 ± 0.22°C to 16.05 ± 0.63°C, respectively (Table 1). The air temperatures were low in the winter season and high in the summer season. Soil temperatures were higher than air temperatures in the winter season and the reverse trend was observed in the summer season. Soil properties of spruce, hemlock, and grassland vegetation types are presented in Fig. 1. Tatachia soils are acidic (pH ranging from 3.40 ± 0.09 to 4.70 ± 0.11), and the pH of deeper horizons were higher than those of the organic layer. The pH of grassland soils was highest, whereas that of hemlock soils was lowest. Moisture contents were high in the organic layer and low in the subsoil. Hemlock soils had the highest moisture content, whereas grassland soils had the lowest. The C_{org} and N_{tot} of spruce and hemlock soils were higher than those of grassland soils. Furthermore, C_{org}, N_{tot}, and C/N ratio were high in the organic layer and low in deeper layers.

Based on the seasonal variation, C_{org} and N_{tot} were high in the spring season and low in the winter season; pH and moisture content were high in the autumn season and low in the summer season. Analysis of variance showed that vegetation types, soil horizons, and sampling seasons had significant to very significant effects on C_{mic} (*p* < 0.01–0.001) and N_{mic} (*p* < 0.001). Vegetation types had significant to very

Table 1 Environmental conditions of spruce, hemlock, and grassland soils in Tatachia forest

Properties	Spruce	Hemlock	Grassland
Characteristic plant	<i>Picea morrisonicola</i>	<i>Tsuga chinensis</i> <i>Y. niitakayamensis</i>	<i>Miscanthus transmorrisonensis</i>
Mean air temperature (°C) ^a			
Autumn	13.35 ± 0.19–13.77 ± 0.21	14.21 ± 0.57–15.03 ± 0.60	14.97 ± 0.20–15.24 ± 0.24
Winter	5.40 ± 0.28–5.83 ± 0.31	6.70 ± 0.20–6.94 ± 0.23	7.50 ± 0.29–7.97 ± 0.33
Spring	14.88 ± 0.61–15.60 ± 0.70	16.32 ± 0.35–16.93 ± 0.40	16.82 ± 0.46–17.21 ± 0.51
Summer	16.14 ± 0.48–16.78 ± 0.52	16.89 ± 0.51–17.51 ± 0.67	20.01 ± 0.53–20.72 ± 0.65
Mean soil temperature (°C) ^a			
Autumn	12.13 ± 0.08–12.31 ± 0.10	12.72 ± 0.13–12.91 ± 0.16	14.57 ± 0.12–14.72 ± 0.13
Winter	6.22 ± 0.22–6.51 ± 0.28	10.52 ± 0.42–11.11 ± 0.63	11.22 ± 0.31–11.53 ± 0.36
Spring	13.01 ± 0.21–13.40 ± 0.30	13.11 ± 0.42–13.82 ± 0.51	13.47 ± 0.44–14.06 ± 0.56
Summer	14.31 ± 0.44–14.75 ± 0.51	14.47 ± 0.40–14.91 ± 0.50	15.23 ± 0.58–16.05 ± 0.63

^a Means of temperature measured during the sampling period (*n* = 3).

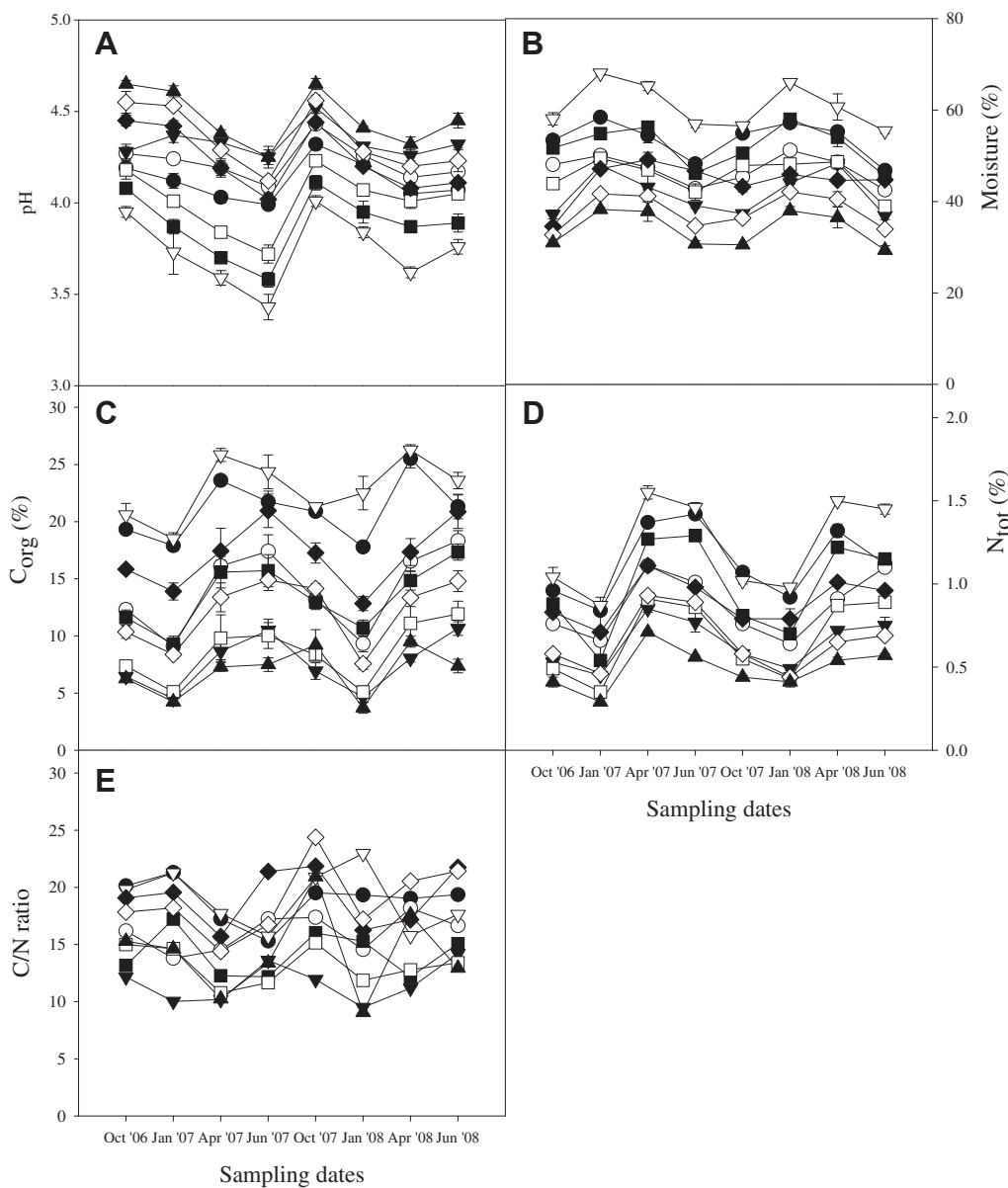


Figure 1. Physicochemical properties of spruce, hemlock, and grassland soils: (A) pH, (B) moisture content, (C) C_{org} , (D) N_{tot} , and (E) C/N ratio. Vertical bars represent the standard deviations ($n = 9$). Symbols used in the figure represent the following: ● = spruce organic layer; ○ = spruce topsoil; ▼ = spruce subsoil; ▽ = hemlock organic layer; ■ = hemlock topsoil; □ = hemlock subsoil; ◆ = grassland organic layer; ◇ = grassland topsoil; ▲ = grassland subsoil. Vertical bars represent the standard deviations ($n = 9$). C_{mic} = microbial biomass carbon; C_{org} = total organic carbon; N_{mic} = microbial biomass nitrogen; N_{tot} = total nitrogen.

significant effects on pH ($p < 0.001$), moisture content ($p < 0.001$), C_{org} ($p < 0.001$), C_{mic}/N_{mic} ($p < 0.001$), and N_{tot} ($p < 0.05$). Soil horizons had significant to very significant effects on moisture content ($p < 0.001$), C_{org} ($p < 0.001$), N_{tot} ($p < 0.001$), C_{org}/N_{tot} ($p < 0.001$), and pH ($p < 0.05$). Sampling seasons had very significant effects on pH ($p < 0.01$) and N_{tot} ($p < 0.01$).

Microbial biomass carbon and nitrogen

The C_{mic} and N_{mic} in three different vegetation types are presented in Fig. 2. In spruce soils, the C_{mic} and N_{mic} ranged from

$380 \pm 10 \mu\text{g/g}$ dry soil to $1320 \pm 25 \mu\text{g/g}$ dry soil and from $74 \pm 5 \mu\text{g/g}$ dry soil to $233 \pm 8 \mu\text{g/g}$ dry soil, respectively. In hemlock soils, the values ranged from $405 \pm 9 \mu\text{g/g}$ dry soil to $1305 \pm 24 \mu\text{g/g}$ dry soil and from $65 \pm 5 \mu\text{g/g}$ dry soil to $275 \pm 10 \mu\text{g/g}$ dry soil, respectively. In grassland soils, the C_{mic} and N_{mic} ranged from $140 \pm 7 \mu\text{g/g}$ dry soil to $855 \pm 15 \mu\text{g/g}$ dry soil and from $48 \pm 4 \mu\text{g/g}$ dry soil to $176 \pm 8 \mu\text{g/g}$ dry soil, respectively. Grassland soils had the lowest C_{mic} and N_{mic} among all tested sites. The C_{mic} and N_{mic} declined substantially with soil depth in all three vegetation types. The organic layer had the highest C_{mic} and N_{mic} , and the values decreased significantly with soil depth ($p < 0.001$). The maximal C_{mic} and N_{mic} were obtained in the spring season and the minimal values in the

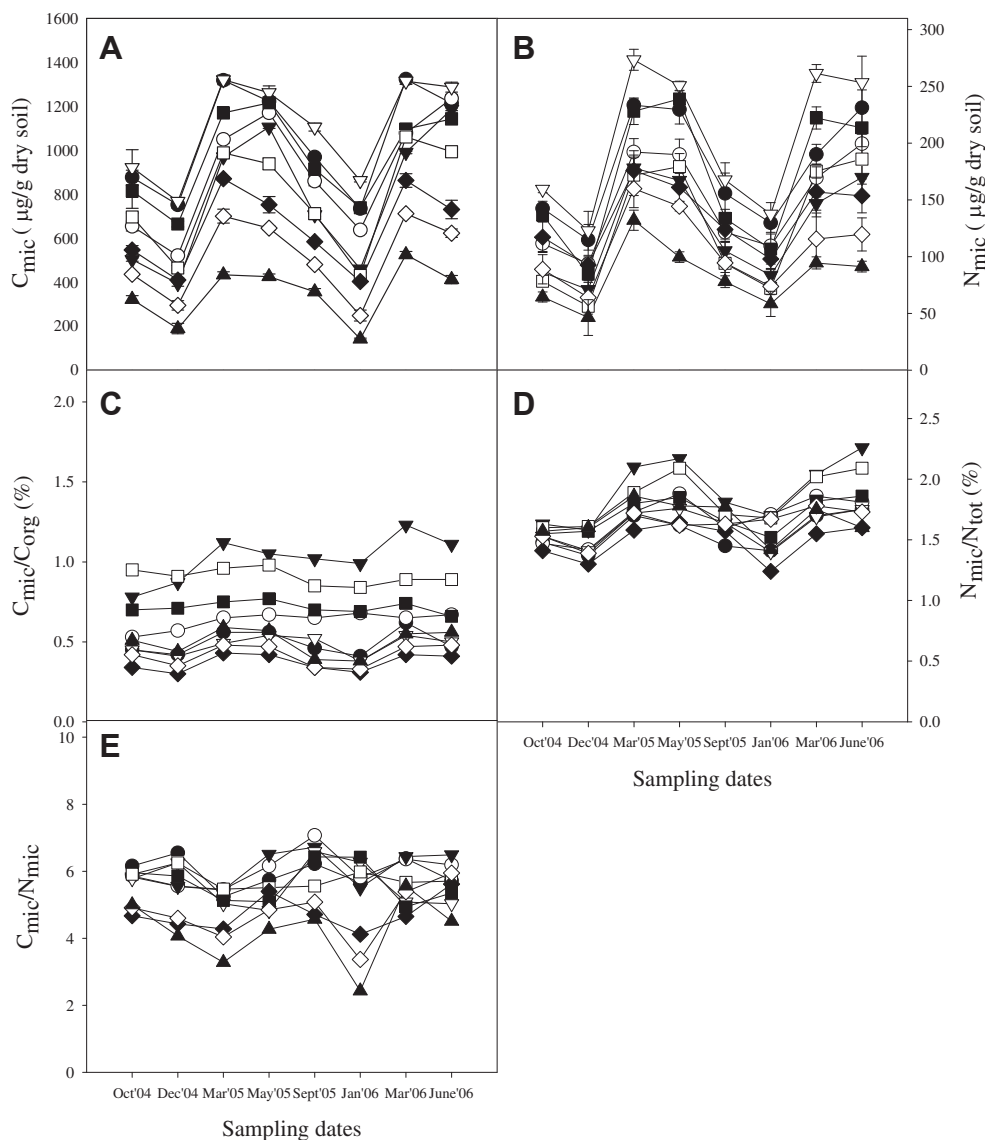


Figure 2. Values of C_{mic} and N_{mic} of spruce, hemlock, and grassland soils: (A) C_{mic} , (B) N_{mic} , (C) C_{mic}/C_{org} (D) N_{mic}/N_{tot} , and (E) C_{mic}/N_{mic} . Vertical bars represent the standard deviations ($n=9$). Symbols used in the figure represent the following: ● = spruce organic layer; ○ = spruce topsoil; ▼ = spruce subsoil; ▽ = hemlock organic layer; ■ = hemlock topsoil; □ = hemlock subsoil; ◆ = grassland organic layer; ◇ = grassland topsoil; ▲ = grassland subsoil. Vertical bars represent the standard deviations ($n=9$). C_{mic} = microbial biomass carbon; C_{org} = total organic carbon; N_{mic} = microbial biomass nitrogen; N_{tot} = total nitrogen.

winter season. The C_{mic}/C_{org} , N_{mic}/N_{tot} , and C_{mic}/N_{mic} ratios increased with soil depth. The highest ratios were observed in spruce soils, followed by for hemlock soils, and the lowest ratios were reported in grassland soils. The C_{mic} and N_{mic} had significant positive correlations with C_{org} ($p < 0.05-0.001$) and N_{tot} ($p < 0.01-0.001$) in all the tested vegetation types.

Microbial populations

Microbial populations of the three tested sites are presented in Fig. 3. Grassland soils had the lowest microbial populations among the tested sites. However, the differences of microbial populations among the three vegetation types are more significant in the organic layer. The organic layer contained the highest populations, which decreased gradually with depth. Microbial populations were highest in the spring season and

lowest in the winter season. Bacterial population was highest among the microbial populations. The ratios of cellulolytic microbes to total microbial populations in organic layers were high due to the roles of carbon cycle.

Discussion

The C_{org} and N_{tot} were high in the organic layer and decreased in the deeper layers, which indicated that the major nutrient pool in the three vegetation types was the organic layer. The C_{mic} and N_{mic} are generally related to the soil organic matter content in forest soils.^{4,7,8} The higher C_{mic} and N_{mic} in the surface soil than in the deeper layers were due to their positive correlations with organic matter content and oxygen availability.³⁶ The C_{mic} and N_{mic} were

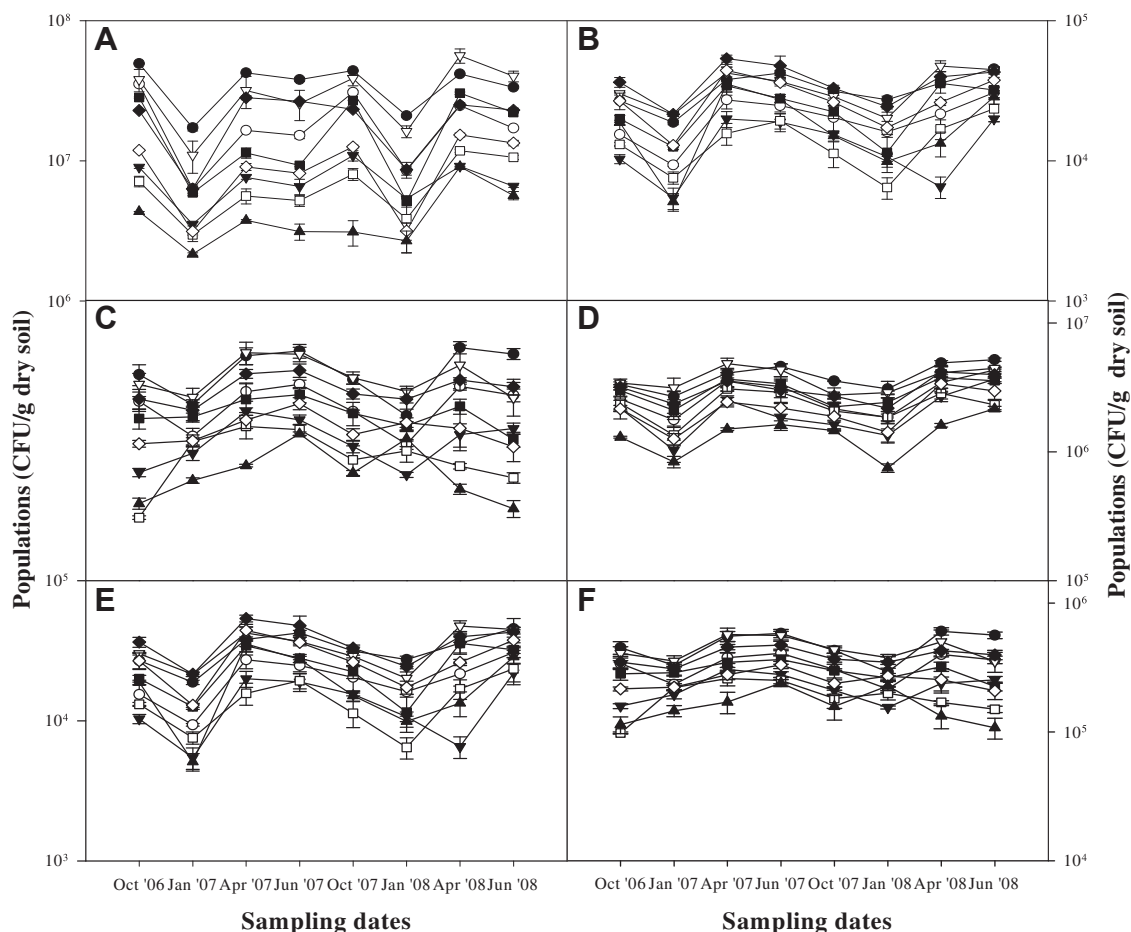


Figure 3. Microbial populations (CFU/g dry soil) of spruce, hemlock, and grassland soils: (A) bacteria, (B) actinomycetes, (C) fungi, (D) cellulolytic microbes, (E) phosphate-solubilizing microbes, and (F) nitrogen-fixing microbes. Symbols used in the figure represent the following: ● = spruce organic layer; ○ = spruce topsoil; ▼ = spruce subsoil; ▽ = hemlock organic layer; ■ = hemlock topsoil; □ = hemlock subsoil; ◆ = grassland organic layer; ◇ = grassland topsoil; ▲ = grassland subsoil. Vertical bars represent the standard deviations ($n = 9$).

higher in the spruce and hemlock soils than in the grassland soils, due to a high tree density and a greater quantity of litters in these soils. In addition, there are large areas of canopy gaps in the grassland patches, and the effective soil layer was shallow due to erosion. Therefore, leaf litter was lowest and soil pH was highest in the three tested soils. Apparently, the differences in substrate quantities between the vegetation types contributed to the differences in C_{mic} and N_{mic} , as indicated by the positive correlations between C_{mic} and N_{mic} and between C_{org} and N_{tot} . Sparling et al.³⁷ had also reported positive correlations between C_{mic} and N_{mic} and between C_{org} and N_{tot} in native forest. The C_{mic} values obtained in the present study were between $140 \pm 7 \mu\text{g/g}$ and $1320 \pm 25 \mu\text{g/g}$. It was within the ranges reported by Vance et al.³⁸ and Henrot and Robertson⁵ in various temperate and tropical forest soils, by Arunachalam and Arunachalam³⁹ in subtropical forests (978–2088 $\mu\text{g/g}$), and by Devi and Yadava⁴⁰ in the mixed oak forest ecosystem (71–1412 $\mu\text{g/g}$). The values of N_{mic} in *Tatchia* forest ranged from $48 \pm 4 \mu\text{g/g}$ to $275 \pm 10 \mu\text{g/g}$, which were also comparable to those in evergreen forest soils (42–242 $\mu\text{g/g}$), broad-leaved deciduous forest soils (132–240 $\mu\text{g/g}$),⁴¹ and forest soils in Germany (317–2116 $\mu\text{g/g}$).⁴² The C_{mic}

and N_{mic} of three tested vegetation types were significantly high in the spring season and low in the winter season. Similar phenomenon was also reported in tropical dry deciduous forests by Saratchandra et al.⁴³ and Singh et al.⁴⁴ Low values of C_{mic} and N_{mic} in the winter season may be due to the low activities of microorganisms and slow rates of decomposition of litters in a dry and cool period.⁴⁰

The C_{mic}/C_{org} ratios in spruce soils (0.4–1.2%) and hemlock soils (0.4–0.96%) were higher than that in grassland soils (0.3–0.6%); the N_{mic}/N_{tot} ratios in spruce soils (1.4–2.3%) and hemlock soils (1.4–2.1%) were also higher than that in grassland soils (1.2–1.7%). These indicated higher C and N immobilization in the spruce and hemlock soils than in the grassland soils. The C_{mic}/C_{org} values obtained in the present study were lower than those of the tropical forest soils (1.5–5.3%)⁴⁵ and tropical wet evergreen forest soils (4–6%),⁴⁶ and similar to those of the subtropical humid forest soils (0.7–1.7%)⁴⁷ and Fushan forest soils (0.3–2.7%).²⁵ The N_{mic}/N_{tot} ratios of the present study were comparable to those of the tropical wet evergreen forest soils (1.3–1.7%),⁴⁶ mixed oak forest soils (0.93–1.8%),⁴¹ and Fushan forest soils (1.5–2.6%);²⁵ but lower than those of the agricultural soils (2–6%),²⁹ forest

soils (3.4–5.9%),⁴⁸ and forest regrowth soils (7.3–8.3%).⁴⁹ These may be due to the low N_{mic} and N_{tot} in the *Tatachia* forest soils, which indicate that the soil is poor in nitrogen. Joergensen et al.⁴² reported that most forest soils had lower N_{mic} than agricultural soils. The C_{mic}/C_{org} and N_{mic}/N_{tot} ratios were high in the spring season and low in the winter season, which indicated high immobilization of C_{mic} and N_{mic} in the spring season.

The C_{mic}/N_{mic} ratio is often used to describe the structure and state of the microbial community. A high C_{mic}/N_{mic} ratio indicates that the microbial biomass contains a high proportion of fungi, whereas a low value suggests that bacteria predominate in the microbial populations.⁴² Paul and Clark⁵⁰ reported that bacterial dominant soil had a C/N ratio between 3 and 5, whereas a C/N ratio between 10 and 15 indicated the dominance of fungi. In the present study, the C_{mic}/N_{mic} ratios of spruce, hemlock, and grassland soils were 5.2–6.5, 4.8–6.6, and 4.1–5.6, respectively, showing the dominance of bacteria.

Microbial growth and metabolism in soils are limited by the availability and types of organic substrates.⁵¹ Among the three tested vegetation types, grassland soils had the lowest microbial populations, which may be due to less C_{org} and N_{tot} . The different plant species have different nutrient demands and produce different qualities and quantities of litter, which affect microbial populations and diversities.⁵² Microbial populations were significantly high in the organic layers and decreased gradually with the depth of the layers in the studied vegetation types. Similar patterns of microbial abundance had been reported in the profiles of agricultural fields⁵³ and Fushan forest soils.²⁵ Miethling et al.⁴⁹ reported that same soil type with different plant species also affected microbial populations.

Conflicts of interest

The authors declare that they have no conflicts of interest related to the subject matter or materials discussed in this article.

Acknowledgments

The authors thank Drs I-Chu Chen, Shu-Hsien Tsai, Selvam Ammaiyappan, and Cheng-Hsiung Chang for their assistances in sampling, Miss Chia-Bei Wei for her technical assistance, and National Science Council of Taiwan for financial support.

References

- Jenkinson DS, Ladd JN. Microbial biomass in soil. Measurement and turnover. In: Paul EA, Ladd JN, editors. *Soil biochemistry*, vol. 5. New York: Marcel Dekker; 1981. pp. 415–71.
- Sun B, Hallett PD, Caul S, Daniell TJ, Hopkins DW. Distribution of soil carbon and microbial biomass in arable soils under different tillage regimes. *Plant Soil* 2010;338:17–25.
- Swift MJ, Heal OW, Anderson JM. *Decomposition in terrestrial ecosystems*. Oxford: Blackwell; 1979.
- Wardle DA. A comparative assessment of factors which influence microbial biomass carbon and nitrogen in soil. *Biol Rev Cambridge Phil Soc* 1992;67:321–58.
- Henrot J, Robertson GP. Vegetation removal in two soils of the humid tropics: effect on microbial biomass. *Soil Biol Biochem* 1994;26:111–6.
- Shrestha RK, Ladha JK, Gami SK. Total and organic soil carbon in cropping systems of Nepal. *Nutrient Cycl Agroecos* 2006;75:257–69.
- Yang SS, Tsai SH, Fan HY, Yang CK, Hung WL, Cho ST. Seasonal variation of microbial ecology in hemlock soil of *Tatachia* mountain, Taiwan. *J Microbiol Immunol Infect* 2006;39:195–205.
- Kujur M, Patel AK. Quantifying the contribution of different soil properties on microbial biomass carbon, nitrogen and phosphorous in dry tropical ecosystem. *Intern J Environ Sci* 2012;2:2272–84.
- Bauhus J, Pare D, Cote L. Effects of tree species, stand age and soil type on soil microbial biomass and its activity in a southern boreal forest. *Soil Biol Biochem* 1998;30:1077–89.
- Liu XM, Li Q, Liang WJ, Jiang Y. Distribution of soil enzyme activities and microbial biomass along a latitudinal gradient in farmlands of Songliao plain, northeast China. *Pedosphere* 2008;18:431–40.
- Nicolardot B, Fauvet G, Cheneby D. Carbon and nitrogen cycling through soil microbial biomass at various temperatures. *Soil Biol Biochem* 1994;26:253–61.
- Gestel MV, Merckx R, Vlassak K. Microbial biomass and activity in soils with fluctuating water contents. *Geoderma* 1993;56:617–26.
- Carter MR. Microbial biomass and mineralizable nitrogen in Solonchic soils: influence of gypsum and lime amendments. *Soil Biol Biochem* 1986;18:531–7.
- Smith JL, Paul EA. The significance of soil microbial biomass estimations. In: Bollag JM, Stotzky G, editors. *Soil biochemistry*, vol. 6. New York: Marcel Dekker; 1990. pp. 357–96.
- Yadav R. Soil organic carbon and soil microbial biomass as affected by restoration measures after 26 years of restoration in mined areas of Doon Valley. *Intern J Environ Sci* 2012;2:1380–5.
- Insam H. Developments in soil microbiology since the mid. *Geoderma* 2001;100:389–402.
- Insam H, Parkinson D, Domsch KH. Influence of macroclimate on soil microbial biomass. *Soil Biol Biochem* 1989;21:211–21.
- Broeckling CD, Broz AK, Bergelson J, Manter DK, Vivanco JM. Root exudates regulate soil fungal community composition and diversity. *Appl Environ Microbiol* 2008;74:738–44.
- Priha O, Smolander A. Microbial biomass and activity in soil and litter under *Pinus sylvestris*, *Picea abies* and *Betula pendula* at originally similar field afforestation sites. *Biol Fertil Soils* 1997;24:45–51.
- Nambu K, Yonebayashi K. Acidic properties of dissolved organic matter leached from organic layers in temperate forests. *Soil Sci Plant Nutr* 1999;45:65–77.
- Rowe EC, Evans CD, Emmett BA, Reynolds B, Helliwell RC, Coull MC, et al. Vegetation type affects the relationship between soil carbon to nitrogen ratio and nitrogen leaching. *Water Air Soil Poll* 2006;177:335–47.
- Sarathchandra SU, Perrot KW, Littler RA. Soil microbial biomass: influence of simulated temperature changes on size, activity and nutrient content. *Soil Biol Biochem* 1989;21:987–93.
- Basiliko N, Moore TR, Lafleur PM, Roulet NT. Seasonal and inter-annual decomposition, microbial biomass, and nitrogen dynamics in a Canadian bog. *Soil Sci* 2005;170:902–12.
- Yang SS, Fan HY, Yang CK, Lin IC. Microbial population of spruce soil in *Tatachia* mountain of Taiwan. *Chemosphere* 2003;52:1489–98.

25. Tsai SH, Selvam A, Yang SS. Microbial diversity of topographical gradient profiles in Fushan forest soils of Taiwan. *Ecol Res* 2007;22:814–24.
26. Kaiser EA, Heinemeyer O. Seasonal variations of soil microbial biomass carbon within the plough layer. *Soil Biol Biochem* 1993;25:1649–56.
27. Vance ED, Brookes PC, Jenkinson DS. An extraction method for measuring soil microbial C. *Soil Biol Biochem* 1987;19:703–7.
28. Wu J, Joergensen RG, Pommerening B, Chaussod R, Brookes PC. Measurement of soil microbial biomass C by fumigation-extraction: an automated procedure. *Soil Biol Biochem* 1990;22:1167–9.
29. Brookes PC, Landman A, Pruden G, Jenkinson DS. Chloroform fumigation and the release of soil nitrogen: a rapid direct extraction method to measure microbial biomass nitrogen in soil. *Soil Biol Biochem* 1985;17:837–42.
30. Yang SS, Chang HL, Wei CB, Lin HC. Reduce waste production with modified Kjeldahl method for nitrogen measurement. *J Biomass Energy Soc China* 1991;10:147–55.
31. Yang CK, Yang SS. Microbial ecology of soils surrounding nuclear and thermal power plants in Taiwan. *Environ Intern* 2001;26:315–22.
32. Mandels M, Reese ET. Induction of cellulase in *Trichoderma viride* as influenced by carbon sources and metals. *J Bacteriol* 1957;73:269–78.
33. Chang CH, Hsieh CY, Yang SS. Effect of cultural media on the phosphate-solubilizing activity of thermotolerant microbes. *J Biomass Energy Soc China* 2001;20:79–90.
34. Nelson DW, Sommers LE. Total carbon, organic carbon and organic matter. In: Page AL, editor. *Methods of soil analysis: part 2. Chemical and microbiological properties*. 2nd ed. Wisconsin: American Society of Agronomy; 1982. pp. 539–80.
35. Stat Sigma. *Sigma Stat user's manual, version 2.0*. San Rafael, CA: Jandel; 1995.
36. Idol TW, Pope PE, Ponder F. Changes in microbial nitrogen across a 100-year chronosequence of upland hardwood forests. *Soil Sci Soc Am J* 2002;66:1662–8.
37. Sparling GP, Hart PBS, August JA, Leslie DM. A comparison of soil and microbial carbon, nitrogen, and phosphorus contents, and macro-aggregate stability of a soil under native forest and after clearance for pastures and plantation forest. *Biol Fertil Soils* 1994;17:91–100.
38. Vance ED, Brookes PC, Jenkinson DS. Microbial biomass measurements in forest soils: the use of the chloroform fumigation incubation method for strongly acid soils. *Soil Biol Biochem* 1987;19:697–702.
39. Arunachalam A, Arunachalam K. Influence of gap size and soil properties on microbial biomass in a subtropical humid forest of north-east India. *Plant Soil* 2000;223:185–93.
40. Devi NB, Yadava PS. Seasonal dynamics in soil microbial biomass C, N and P in a mixed-oak forest ecosystem of Manipur, northeast India. *Appl Soil Ecol* 2006;31:220–7.
41. Diaz-Ravina M, Carballas T, Acea MJ. Microbial biomass and metabolic activity in four acid soils. *Soil Biol Biochem* 1988;20:817–23.
42. Joergensen RG, Anderson TH, Wolters T. Carbon and nitrogen relationships in the microbial biomass of soils in beech (*Fagus sylvatica*) forests. *Biol Fertil Soils* 1995;19:141–7.
43. Saratchandra SU, Perrot KW, Upsdell MP. Microbiological and biochemical characteristics of a range of New Zealand soils under established pasture. *Soil Biol Biochem* 1984;16:177–83.
44. Singh JS, Raghubanshi AS, Singh RS, Srivastava SC. Microbial biomass acts as a source of plant nutrients in dry tropical forest and savanna. *Nature* 1989;338:499–500.
45. Luizao RCC, Bonde TA, Rosswall T. Seasonal variation of soil microbial biomass—the effect of clear felling in a tropical rain forest and establishment of pasture in the Central Amazon. *Soil Biol Biochem* 1992;24:805–13.
46. Barbhuiya AR, Arunachalam A, Pandey HN, Arunachalam K, Khan ML, Nath PC. Dynamics of soil microbial biomass C, N and P in disturbed and undisturbed stands of a tropical wet-evergreen forest. *Eur J Soil Biol* 2005;40:113–21.
47. Maithani K, Tripathi RS, Arunachalam A, Pandey HN. Seasonal dynamics of microbial biomass C, N and P during regrowth of a disturbed subtropical humid forest in northeast India. *Appl Soil Ecol* 1996;4:31–7.
48. Martikainen PJ, Palojarvi A. Evaluation of the fumigation extraction method for determination of microbial C and N in a range of forest soils. *Soil Biol Biochem* 1990;27:797–802.
49. Miethling R, Wieland G, Backhaus H, Tebbe CC. Variation of microbial rhizosphere communities in response to crop species, soil origin and inoculation with *Sinorhizobium meliloti* L33. *Microb Ecol* 2000;40:43–56.
50. Paul EA, Clark FE. *Soil microbiology and biochemistry*. 2nd ed. London: Academic; 1996. pp. 129–55.
51. Myers RT, Zak DR, White DC, Peacock A. Landscape-level patterns of microbial community composition and substrate use in upland forest ecosystems. *Soil Sci Soc Am J* 2001;65:359–67.
52. Bezemer TM, Lawson CS, Hedlund K, Edwards AS, Brooks AJ, Igual JM, et al. Plant species and functional group effects on abiotic and microbial soil properties and plant-soil feedback responses in two grasslands. *J Ecol* 2006;94:893–904.
53. Taylor JP, Wilson B, Mills MS, Burns RG. Comparison of microbial numbers and enzymatic activities in surface soils and subsoils using various techniques. *Soil Biol Biochem* 2002;34:387–401.